



Inferring Total Species Richness and the Exhaustive Hierarchical Structuring of Species Abundances in Tropical Sea-stars Communities (Asteroidea), Using Numerical Extrapolation of Partial Inventories

Jean Béguinot^{1*}

¹*Société Histoire Naturelle-Bourgogne Nature, 7 bvd H.P. Schneider 71200, Le Creusot, France.*

Author's contribution

The sole author designed, analysed, interpreted and prepared the manuscript.

Article Information

DOI: 10.9734/AJEE/2018/46272

Editor(s):

(1) Dr. Daniele De Wrachien, Professor, Agricultural Hydraulics at the Department of Agricultural and Environmental Sciences, State University of Milan, Italy.

Reviewers:

(1) Manoel Fernando Demétrio, Universidade Federal da Grande Dourados, Dourados, Brazil.

(2) Retno Hartati, Diponegoro University, Indonesia.

Complete Peer review History: <http://www.sdiarticle3.com/review-history/46272>

Original Research Article

Received 09 October 2018
Accepted 28 December 2018
Published 15 January 2019

ABSTRACT

Even when remaining substantially incomplete, the partial inventory of a species assemblage can provide much more information than could be expected first. Indeed, retrieving this information is made possible by applying a rigorous procedure of numerical extrapolation to the partial inventory. This numerical extrapolation will deliver reliable estimates of the number of the still unrecorded species and, furthermore, of the distribution of abundances of these unrecorded species. As a result, the full range of the Species Abundance Distribution is finally made available, despite dealing with data from partial sampling only. In turn, this allows to address a series of descriptive and functional aspects of the internal organization of species assemblages, which otherwise would have required disposing of exhaustive samplings. The latter being, however, often impossible to achieve in practice. Fortunately, mathematic and algorithmic basis for a reliable numerical extrapolation of incomplete samplings have been developed recently, so that partial inventories no longer remain an obstacle to gain access to the true (total) species richness and the full-range pattern of hierarchical structuring of species abundances.

This approach is applied here to the previously reported partial samplings of two communities of tropical sea-stars associated to coral-reefs in the Central South China Sea. Among the main new

*Corresponding author: E-mail: jean-beguिनot@orange.fr;

results derived from the numerical extrapolations of these partial samplings, the following are highlighted:

- the extrapolated true (total) species richness of each of the two studied communities largely exceed the recorded figures, thereby confirming the limited completeness of these partial samplings: 53% and 67% completeness only;
- once properly completed by numerical extrapolation, the Species Abundance Distributions of both communities prove best fitting the “log-normal” than the “log-series” model, thereby suggesting that many independent factors (rather than only one dominant factor) contribute together to the hierarchical structuring of species abundances within these sea-stars communities;
- the intensity of the process of hierarchical structuration of species abundances proves being close to what would be obtained for the corresponding “broken-stick” model, which means a rather moderate level of structuration intensity, as compared to the range of values typically obtained for other kinds of marine invertebrates.

Keywords: *Echinodermata; starfish; coral reef; species richness; diversity; rank abundance distribution; evenness; incomplete inventory; Spratly Islands, Malaysia.*

1. INTRODUCTION

Total species richness, taxonomic composition and the hierarchical structuring of species abundances are three key features that appropriately characterize species communities in the wild. Addressing properly these key features obviously requires disposing of exhaustive inventories. Yet, inventories remaining substantially *incomplete* are common in practice and even doomed to become still more frequent with the inevitable generalization of “rapid assessments” and “quick surveys”. This is especially true when having to deal with species-rich communities of invertebrates which often include a lot of rare, hard-to-detect species. Such incomplete inventories prevent deriving reliable inferences and, thus, may often lead to erroneous interpretations regarding the key aspects of species communities evoked above [1–3].

Fortunately, a reliable procedure of *numerical extrapolation* of partial inventories has been developed recently, which can overcome these difficulties and is able to provide least-biased estimations of:

- (i) The number of those species remained undetected and, still further,
- (ii) The respective abundances of each of these undetected species.

Thereby, reliable inferences can finally be derived regarding (i) the *true* (total) species richness and (ii) the *completed* distribution of species abundances, i.e. including the set of the still unrecorded species. Only the taxonomic identities of these undetected species escape, of course, to any attempt to highlight them. In turn, once numerically *completed* (and *only* when it is

so: [4]), the Species Abundance Distribution (“S.A.D.”) can then provide synthetic pieces of information about the process (either deterministic or stochastic) that drives the hierarchical structuring of species abundances within community [5–9]. Accordingly, some light can thus be shed, both qualitatively and quantitatively, on the biological and ecological determinants of the internal structuration among species within community.

Although no further details may be extracted from this synthetic overview, the latter has, yet, the advantage of being straightforward, as it does not require the long and tedious analytical approach that would be required otherwise to go deeper in the details of structuring process. As such, this synthetic overview can serve as a convenient preliminary approach.

Hereafter, I report and discuss the results from the numerical extrapolation of the partial inventories of two sea-stars communities (Echinodermata: Asteroidea) associated to coral reefs surrounding the small closely neighboring islets of Terumbu Siput (“Erica reef”) and Terumbu Peninjau (“Investigator shoal”), parts of Spratly Islands, in the Central South China Sea, off the Malaysian coast. Such marine ecosystems, in tropical shallow waters, are of major interest to ecologists and conservationists, as they are considered as embodying remarkably high levels of biological complexity [10 – 12].

2. MATERIALS AND METHODS

2.1 Materials

Coral reefs surrounding small islands dispersed in the Central South China Sea are home to a rich sea-stars fauna, reported as counting overall

no less than 230 species [13]. Yet, inventories at the local scale of sea-stars communities in China Sea, additionally listing the abundances of the recorded species, remain very scarce at the *local scale*. A recent report by Kwang et al. [12], however, opportunely provides such a series of local inventories of sea-stars from the Archipelago of Beting Patinggi Ali to Pulau Layang-Layang in the Malaysian waters of Central South China Sea. Yet, the high proportion of singletons (species detected only once during sampling) that subsist in these inventories strongly suggests that samplings remain substantially incomplete [14–16], thus requiring *numerical extrapolation* to take full advantages of the as-recorded data and avoid the risk of erroneous inference.

Hereafter, I focus on two of these investigated sea-stars communities, those having the highest number of recorded species (16 and 9 species) respectively. These two communities are located in two small neighboring coral islets: *Terumbu Siput* (“Erica reef”) and *Terumbu Peninjau* (“Investigator shoal”), both being part of “Spratly Islands”, off the coasts of the Philippines, Malaysia, and southern Vietnam. Further information regarding the precise locations, the environment and the practical details of the sampling procedure are provided in open reference [12].

2.2 Numerical Extrapolation Procedure

2.2.1 Total species richness

The least-biased estimation of the number of still undetected species after partial sampling and the resulting least-biased estimation of the true, total species richness of a partially sampled community are derived according to the procedure defined in [17,18] and briefly summarized in Appendix 1, on the basis of the numbers f_x of species recorded x -times during the two partial samplings: Figs. A.1 and A.2).

2.2.2 Completed species abundance distribution

To accurately exploit its full potential, the as-recorded Species Abundance Distribution (“S.A.D.”) requires [19,20]:

- first, to be *corrected* for the statistical sampling bias resulting from the finite size of samplings;
- second, but still more importantly, to be *completed* by numerical extrapolation,

insofar as the sampling is suspected to be incomplete (as revealed by the subsistence of several singletons).

After being *corrected* and *completed* by *numerical extrapolation*, the S.A.D.:

- not only provides an overview of both the true species richness of the sampled community and the diversity of the respective abundances of member-species
- but, also, can help addressing important questions regarding (i) the kind of causes that determine the hierarchical structuration of species abundances, (ii) the resulting degree of abundance unevenness and (iii) the genuine intensity of the hierarchical structuring process (which by no means identifies to the degree of unevenness, contrary to a commonly held opinion).

The appropriate procedures of (i) correction and (ii) numerical extrapolation of the “S.A.D.” beyond its as-recorded part, are described in details in reference [20] and briefly summarized in Appendix 2. Also, a concrete example of implementation of the procedure is commented in details in reference [21].

Classically, the “S.A.D.” is graphically presented according to the so-called “Ranked Abundance Distribution” (also known as “Whittaker plot”), according to which the (log-transformed) abundances a_i are plotted against the rank i of species, the latter being ordered by decreasing values of abundance (with, thus, a_1 and a_{S_t} respectively standing for the highest and the lowest abundances in a community having S_t species).

2.2.3 Species abundance structuration: the apparent pattern of abundance unevenness

The “S.A.D.” (being either exhaustive or completed by numerical extrapolation) conveys all the relevant quantitative data required to address the internal organization of member-species within a local community, especially the hierarchical structuration of species abundances. In particular, it is always advisable to use such species-abundance plots to quantify the degree of evenness or, more suggestively, the degree of *unevenness* of species abundances [22]. Indeed, following [23], it is the degree of *unevenness* – rather than evenness itself – that should be

preferred to address properly the hierarchical structuring of species abundances within communities. Optionally, the “S.A.D.” may be synthetically reduced to its two major descriptors: the *total species richness* S_t and the degree U of *abundance unevenness*.

According to the aforementioned, classical mode of representation of the “S.A.D.”, it goes natural to quantify the degree U of abundance unevenness as the average of the decreasing slope of the log-transformed abundance along the whole range of the abundance distribution, as already proposed by Grzès [24], that is:

$$\begin{aligned} U &= [\log_{10}(a_1) - \log_{10}(a_{S_t})] / (S_t - 1) \\ U &= [\log_{10}(a_1 / a_{S_t})] / (S_t - 1) \end{aligned} \quad (1)$$

2.2.4 Species abundance structuration: the origin and intensity of the underlying process

Beyond the mere description of the pattern of hierarchical structuration, quantified by the degree of unevenness U , the complete “S.A.D.” can help addressing several important questions regarding (i) the kind of *mechanism* involved in the process driving the hierarchical structuration of abundances and (ii) the *intensity* of this structuring process.

As regards the kind of *mechanism* involved, it is appropriate to distinguish between two major alternative hypotheses: schematically, the hierarchical structuration of abundances may result either (i) from the major contribution of *one strongly predominant* factor or (ii) from the combined contributions of *many mutually independent factors* acting together. This can be tested by checking the conformity of the “S.A.D.” to either the *log-series* model or the *log-normal* model respectively [5,25–28].

Now, as regards the genuine *intensity* of the structuring process, it is first necessary to remind that the degree of unevenness U does not univocally mirror the intensity of the structuring process, since it is *also* mathematically dependent (negatively) upon the species richness S_t [29,30,31]; see also Appendix 3.

One possible solution to cancel this mathematical influence is to compare the slope of the “S.A.D.” with the slope of a theoretical distribution involving a *constant* structuring process, remaining *strictly independent* of the species richness. The “broken-stick” distribution meets precisely this requirement [32].

Accordingly, an index “ I_{str} ”, attempting to disentangle the biologically relevant part of the unevenness - that is *independent of species richness* - can be defined by standardizing the degree of unevenness U of the “S.A.D.” to the degree of unevenness U' of the corresponding “broken-stick” distribution, computed for the same species richness S_t [33-35]. That is:

$$\begin{aligned} I_{str} &= U/U' \\ &= [\log_{10}(a_1/a_{S_t})/(S_t-1)]/[\log_{10}(a'_1/a'_{S_t})/(S_t-1)] \end{aligned}$$

that is finally:

$$I_{str} = \log_{10}(a_1 / a_{S_t}) / \log_{10}(a'_1 / a'_{S_t}) \quad (2)$$

with a_1 and a_{S_t} standing for the highest and the lowest abundances in the studied community and a'_1 and a'_{S_t} standing for the highest and the lowest abundances in the corresponding “broken-stick” distribution computed for the same species richness S_t .

As the *invariable type* of structuring process involved in the “broken-stick” distribution is the process of randomized allocation of abundance values to species [32], the index I_{str} highlights the *intensity* of the *structuring process* in the considered community by comparison to the intensity corresponding to this stochastic process, taken as a meaningful reference.

It is possible to continue the analysis even further, in the same direction. Beyond comparing the average slopes of the actual “S.A.D.” to the corresponding “broken-stick” model, it is additionally informative to operate the comparison separately for each of the two extremal points that support the slopes – i.e. the maximum and the minimum abundances, a_1 and a_{S_t} . Thus, the ratio $A_1 = (a_1 / a'_1)$ mirrors the intensity of those kind(s) of factors which control the abundances of the more frequent species (especially the abundance a_1 of the most frequent one) while the ratio $A_{S_t} = (a_{S_t} / a'_{S_t})$ mirrors the intensity of those (presumably different) kind(s) of factors which control the abundances of the less frequent species (especially a_{S_t} for the least abundant species).

Split apart that way, the comparison with the “broken-stick” model (synthetically expressed by the index I_{str}), highlights even better the underlying structuring *process* and its *genuine* intensity, beyond the immediately apparent pattern of abundance unevenness U . This is briefly detailed in the following section.

2.2.5 Analyzing the determinants of the species abundance structuration

Focused on quantitative terms, the “S.A.D.” can be synthesized by three main descriptive parameters, S_t , U , a_{St} (the fourth descriptive parameter, the higher abundance a_1 , being entirely defined by the three preceding parameters through equation (1)). Now, which “determinants” are actually constraining the values of S_t , a_1 and a_{St} ?

Let consider first the case where the community has hypothetically reached its saturation level in term of species richness S_t , i.e. no more species may be added in the community without causing extinction(s) [36]. This means that both the abundance unevenness U and the lowest abundance a_{St} have reached their respective minimal threshold values that would necessarily be crossed over if any additional colonization would succeed in increasing species richness beyond the saturation level. Accordingly, in case of saturation, the species richness S_t is determined internally, the parameters U and a_{St} playing the role of determinants.

Now, it is widely admitted that the saturation in species within communities is very uncommon [36,37,38], so that non-saturation should be hypothesized first. In non-saturated communities, the species richness S_t is no more determined internally but externally, the determinant being the limitation in the colonizing flux of those species having characteristics compatible with the habitat. A limitation which depends on both the dispersal abilities and the richness of the regional stock of species [37,38,39]. And, by contrast with an hypothetically saturated community, here, the parameters U and a_{St} are no more involved as the determinants of S_t . On the contrary, in this situation of limited availability of new colonizing species, it is the resulting limitation of species richness S_t which

contributes to constrain the possible ranges of values of the triplet of parameters U , a_{St} and a_1 , through equation (1). And, finally, it is I_{str} which is ultimately constrained, according to equation (2), as are the ratios $A_1 = (a_1/a'_1)$ and $A_{St} = (a_{St}/a'_{St})$.

3. RESULTS

3.1 Estimation of the Total Species Richness of Each Sea-Stars Community

Based on the numbers f_x of species recorded x -times (with $x = 1$ to 5) at the end of the partial samplings (Figs. A1 and A2 in Appendix 1), the selected, least-biased estimators of the number of undetected species are respectively Jackknife at order 3 and Jackknife at order 5 for the communities at Terumbu Siput and Terumbu Peninjau (see selective key in Appendix 1). The corresponding least-biased estimations of (i) the number Δ of undetected species, (ii) the resulting least-biased estimation of the total species richness S_t of the sampled communities and (iii) the level of completeness of the partial samplings, R_0/St , are provided in Table 1.

Due to the relatively low level of achieved sampling completeness (67% and 53%), further additional sampling could, alternatively, be considered of interest. In this perspective, the least-biased extrapolation of the Species Accumulation Curve can provide useful predictive information regarding the additional sampling efforts that would be required to obtain any desirable increase in sampling completeness. As an example, the expected increase in the number of detected species, $R(N)$, as a function of growing sampling size N , is given in Figs. 1 and 2 for the sampling at Terumbu Siput and in Figs. 3 and 4 for the sampling at Terumbu peninjau.

Table 1. Numerical characteristics of the Sea-Stars communities associated to coral reefs at Terumbu Siput, and Terumbu Peninjau: the sampling-size N_0 , the number of detected species $R_0 (= R(N_0))$, the selected, least-biased estimator, the estimated number Δ of undetected species, the resulting evaluation of the total species richness $S_t = R_0 + \Delta$ and the level of sampling completeness R_0/St

Site	N_0	R_0	selected estimator	Δ	S_t	R_0/St
Terumbu Siput	52	16	Jack - 3	8	24	67 %
Terumbu Peninjau	26	9	Jack - 5	8	17	53 %

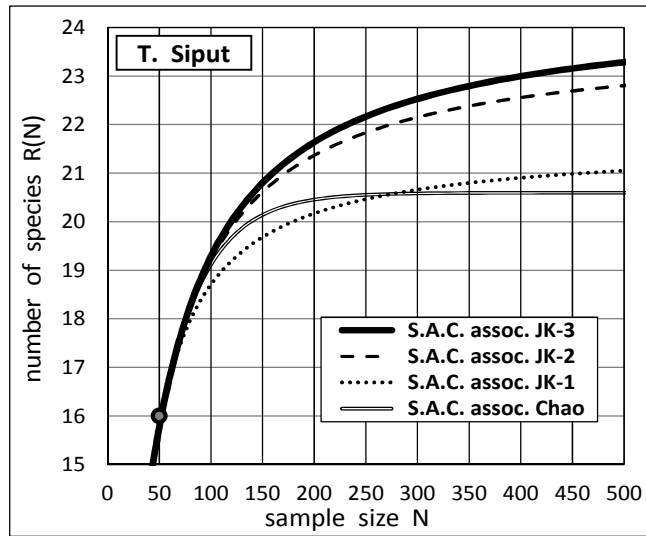


Fig. 1. Extrapolated part of the species accumulation curve (S.A.C.) for the community at *Terembu siput*, accounting for the increase of the number of detected species $R(N)$ as a function of growing sample size N , beyond the actually achieved sampling ($N_0 = 52$, $R(N_0) = 16$). Here, the selected, least-biased, nonparametric estimator of the number of undetected species is Jackknife-3, leading to a total species richness $S_t = 16 + 8 = 24$. The associated, least-biased extrapolation of the S.A.C., $R(N)$, is plotted as the coarse solid line. Also plotted, for comparison, are the extrapolations of the S.A.C. associated to other, non-selected (as being more biased) estimators: Jackknife-2, Jackknife- 1, Chao 1. The comparison highlights the practical importance of selecting the *least-biased* estimator

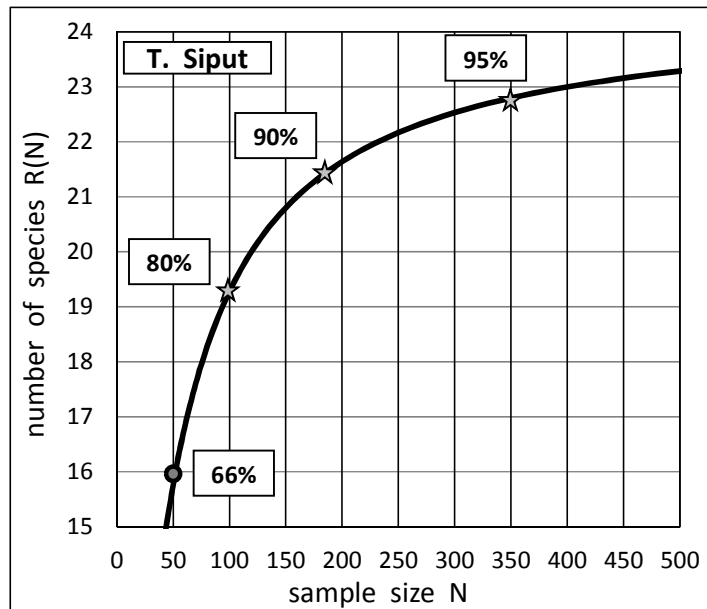


Fig. 2. Extrapolated part of the species accumulation curve for the community at *Terembu siput*, associated to the selected, least-biased, nonparametric estimator (here Jackknife-3). In practice, the least-biased extrapolation of the Species Accumulation Curve allows to predict the expected additional sampling effort required to reach higher levels of sampling completeness (for example, the sample sizes required to reach 80%, 90% and 95% completeness would be around $N = 100, 180, 350$ respectively)

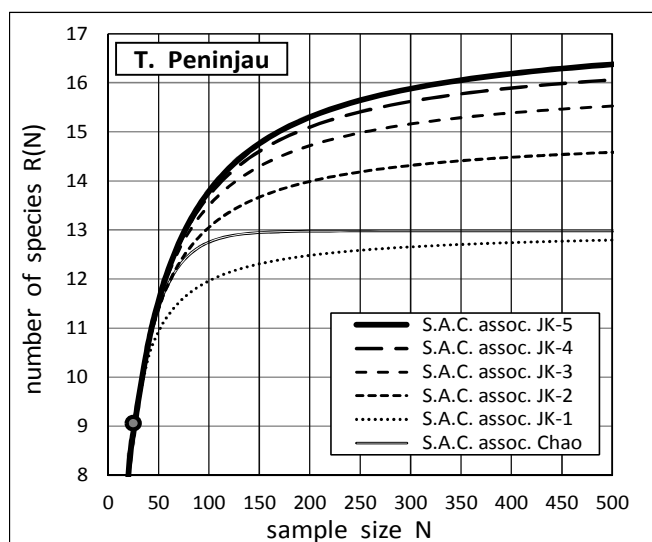


Fig. 3. Extrapolated part of the species accumulation curve for the community at *Terumbu peninjau*, accounting for the increase of the number of detected species $R(N)$ as a function of growing sample size N , beyond the actually achieved sampling ($N_0 = 26$, $R(N_0) = 9$). Here, the selected, least-biased, nonparametric estimator of the number of undetected species is Jackknife-5, leading to a total species richness $S_t = 9 + 8 = 17$. The associated, least-biased extrapolation of the species accumulation curve, $R(N)$, is plotted as the coarse solid line. Also plotted, for comparison, are the extrapolations of the S.A.C. associated to other, here non-selected (more biased) estimators: Jackknife 4, Jackknife 3, Jackknife-2, Jackknife- 1, Chao1. The comparison highlights the practical importance of selecting the *least-biased* estimator

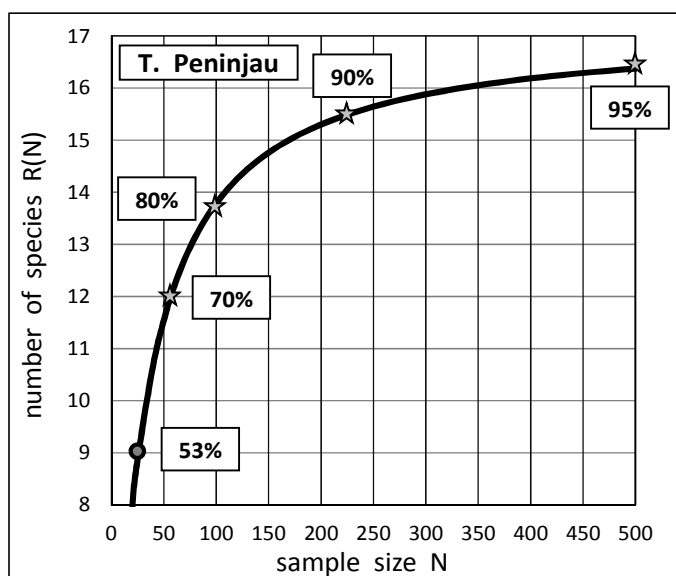


Fig. 4. Extrapolated part of the species accumulation curve for the community at *Terumbu peninjau*, associated to the selected, least-biased, nonparametric estimator (here Jackknife-5). In practice, the least-biased extrapolation of the species accumulation curve allows to predict the expected additional sampling effort required to reach higher levels of sampling completeness (for example, the sample sizes required to reach 70%, 80%, 90% and 95% completeness would be around $N = 60, 100, 220, 500$ respectively)

3.2 Correction and Extrapolation of the Species Abundance Distributions

The bias-corrected and extrapolated Species Abundance Distributions are plotted in Fig. 5 (recorded part: ranks $i = 1$ to 16; extrapolated part from rank 17 to 24) for the community at Terembu Siput and in Fig. 6 (recorded part: Ranks $i = 1$ to 9; extrapolated part from rank 10 to 17) for the community at Terembu Peninjau. Figs. 5 and 6 thus provide the *entire development* of the Species Abundance Distribution. Note that the *extrapolated* part of the distribution of abundances has no less importance than the *recorded* part since more or less rare species may have as equal ecological importance as more common ones [40 – 49]. Fig. 7 allows to compare directly the Species Abundance Distributions of the two communities.

3.3 Qualitative and Quantitative Characterizations of the Hierarchical Structuring of Species Abundances

3.3.1 The type of process likely involved in the hierarchical structuring of species abundances

In order to investigate which kind of mechanism is involved in the structuring process at work in these sea-stars communities, two classical models of abundance distribution – the “log-normal” distribution and the “log-series” distribution – were tentatively fitted to the completed Species Abundance Distributions provided at Figs. 5 and 6. As shown in Figs. 8 to 11, a fairly good fit is obtained with the *log-normal* model for both studied communities, while the accordance with the *log-series* model is comparatively unsatisfactory.

As regards the taxonomically identified species (i.e. those species that were actually

encountered and identified during partial samplings), their compared ranks along the hierarchy of abundances are highlighted in Figs. 12 and 13. Not surprisingly, the three most common species (labelled a, b, c in Figures) are the same ones in both communities: *Linckia laevigata*, *Linckia multifora*, *Culcita novaeguineae*. On the contrary, the identities of those species having lesser abundances quite largely differ between both communities.

3.3.2 The intensity of the hierarchical structuring process

According to the definitions provided in Methods section:

- the degree, U , of unevenness of species abundance distribution resulting from the process driving the hierarchical structuration of abundances is computed according to equation (1);
- the genuine intensity, I_{str} , of the process driving the hierarchical structuration of abundances is relevantly appreciated by comparing the “S.A.D.” of the studied community to the corresponding “broken-stick” model, computed for the same species richness. Figs. 14 and 15 allows this comparison, from which the genuine intensity I_{str} of the structuring process is derived according to equation (2).

The corresponding results are summarized in Table 2 which highlights (i) the true total species richness S_t , (ii) the ratio a_1 / a_{St} between the abundances of the commonest and the rarest species, (iii) the degree of unevenness of species abundances U and, finally, (iv) the genuine intensity I_{str} of the process driving the hierarchical structuration of species abundances. Note that the parameters U and I_{str} accounts for

Table 2. A synthetic summary of the main quantitative features of the hierarchical organization of species abundances within community, as derived from each numerically completed “S.A.D.” : (i) the total species richness S_t of the community ; (ii) the relative abundances a_1 and a_{St} of the most and the least abundant species (species rank 1 and S_t) ; (iii) the unevenness of abundances in the community: $U = \log(a_1/a_{St})/(S_t - 1)$; (iv) the unevenness of abundances in the corresponding “broken-stick” distribution: $U' = \log(a'_1/a'_{St})/(S_t - 1)$, (v) the genuine intensity of the structuring process $I_{str} = U/U'$ and, at last, the ratios $A_1 = (a_1 / a'_1)$ and $A_{St} = (a_{St} / a'_{St})$

Sites	S_t	a_1	a_{St}	a_1/a_{St}	a'_1	a'_{St}	U	U'	I_{str}	A_1	A_{St}
T. Siput	24	0.14422	0.00157	94	0.15733	0.00174	0.0854	0.0851	1.003	0.92	0.90
T. Peninjau	17	0.19341	0.00197	98	0.20233	0.00346	0.1246	0.1104	1.128	0.96	0.57

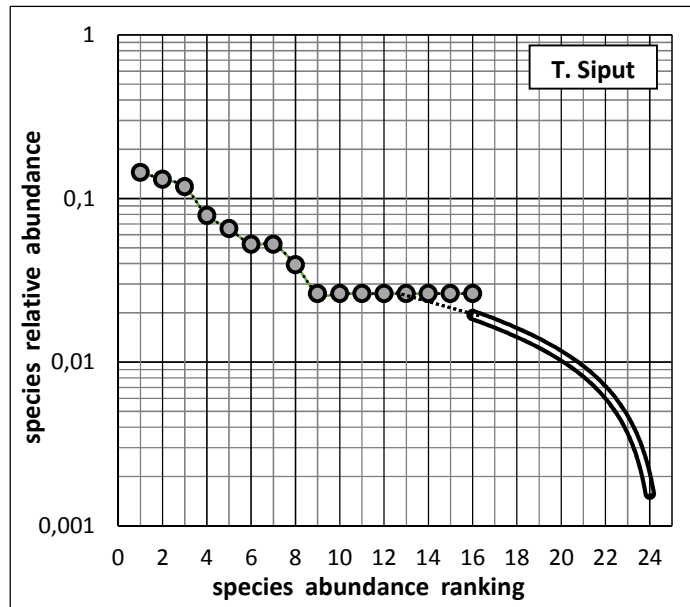


Fig. 5. The completed species abundance distribution for the community at *Terembu Siput*, including the correction of the recorded part (involving the 16 detected species: grey discs) and the least-biased extrapolation of the unrecorded part (involving the 8 undetected species: Coarse double line). Note logarithmic scale for relative abundances, a classical convention of graphical representation

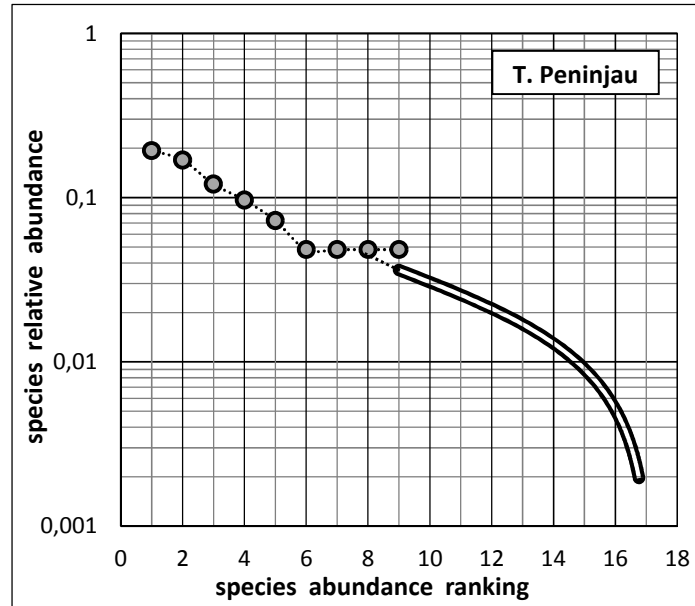


Fig. 6. The completed species abundance distribution for the community at *Terembu Peninjau*, including the correction of the recorded part (involving the 9 detected species: Grey discs) and the least-biased extrapolation of the unrecorded part (involving the 8 undetected species: Coarse double line). Note logarithmic scale for relative abundances, a classical convention of graphical representation

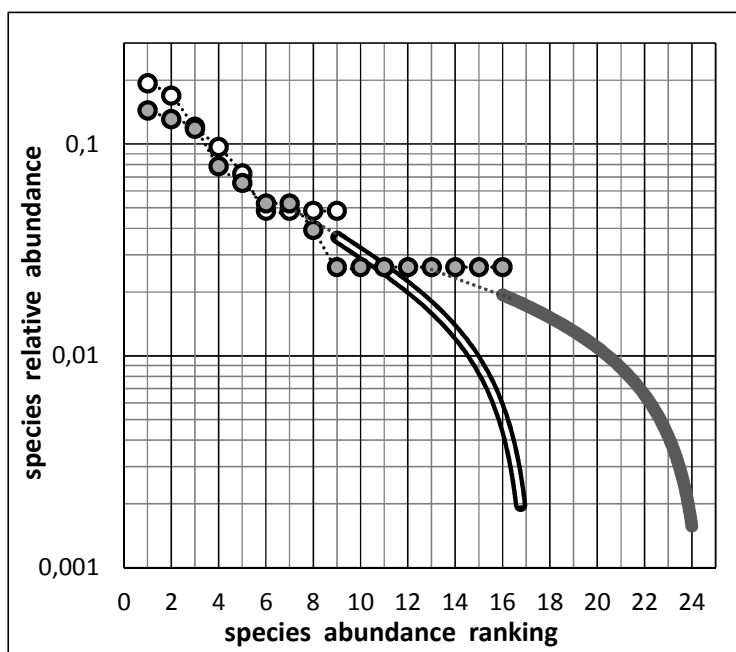


Fig. 7. Direct comparison between the completed species abundance distributions of both communities - grey figures: *Terumbu Siput*; white figures: *Terumbu Peninjau*

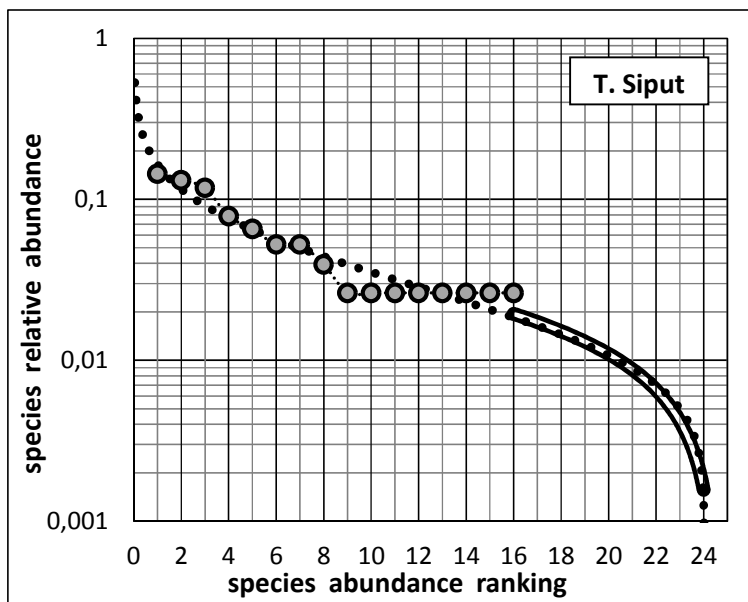


Fig. 8. The classical “log-normal” model (sigmoid dotted line) fitted to the completed Species Abundance Distribution of the community at *Terumbu Siput*.

two complementary aspects of the hierarchical structuration of species abundances: while U quantifies the apparent *pattern* of species abundance structuration, I_{str} highlights the genuine intensity of the underlying *process* driving this structuration. Being understood that,

in quantitative terms, the unevenness pattern is far from faithfully reflecting the structuring process itself, as already emphasized.

The ratios $A_1 = (a_1/a'_1)$ and $A_{St} = (a_{St}/a'_{St})$ (which mirror the intensity of those factors which control

the abundances a_1 and a_{St} of the most and the least frequent species: see section Methods) are derived accordingly: $A_1 = 0.92$, $A_{St} = 0.90$ at

Terumbu Siput and $A_1 = 0.96$, $A_{St} = 0.57$ at Terumbu Peninjau (Figs. 16 and 17).

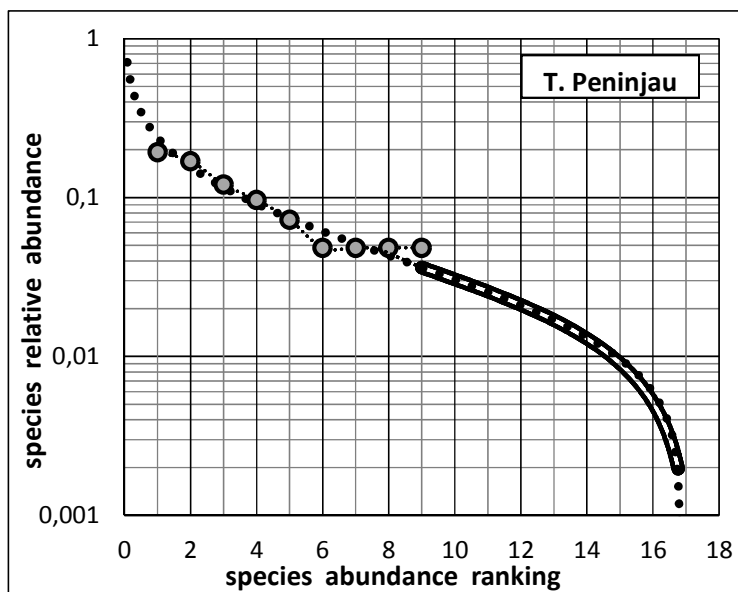


Fig. 9. The classical “log-normal” model (sigmoid dotted line) fitted to the completed species abundance distribution of the community at *Terumbu Peninjau*

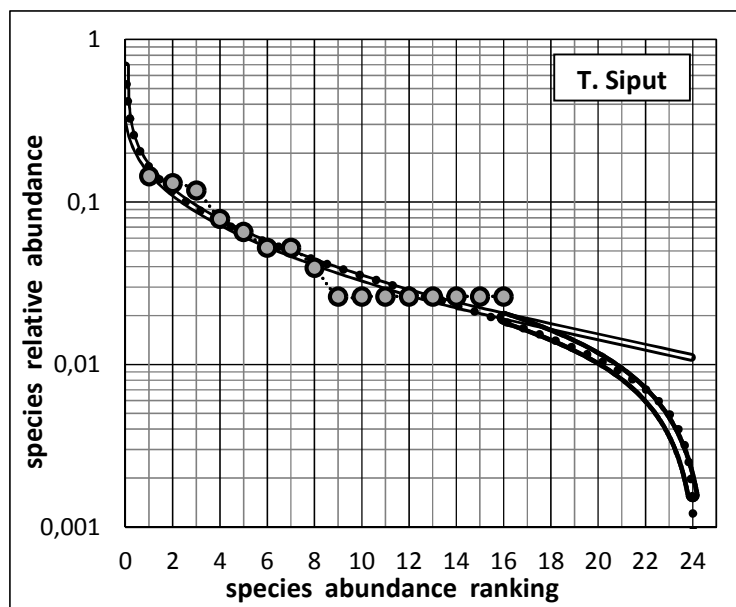


Fig. 10. The two classical models: “log-normal” (sigmoid dotted line) and “log-series” (fine double line) compared to the Species Abundance Distribution of the community at *Terumbu Siput*. Best fit is clearly obtained with the “log-normal” distribution. Note that considering the *recorded part* of the Species Abundance Distribution *only* would have led to the opposite conclusion: a “J” shaped model such as the “log-series” model would have fit the recorded part best than a “sigmoid” shaped model such as the “log-normal” model

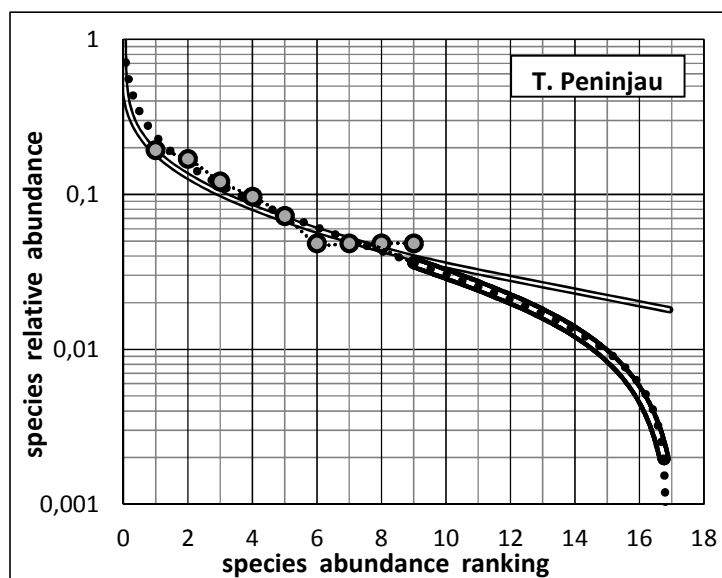
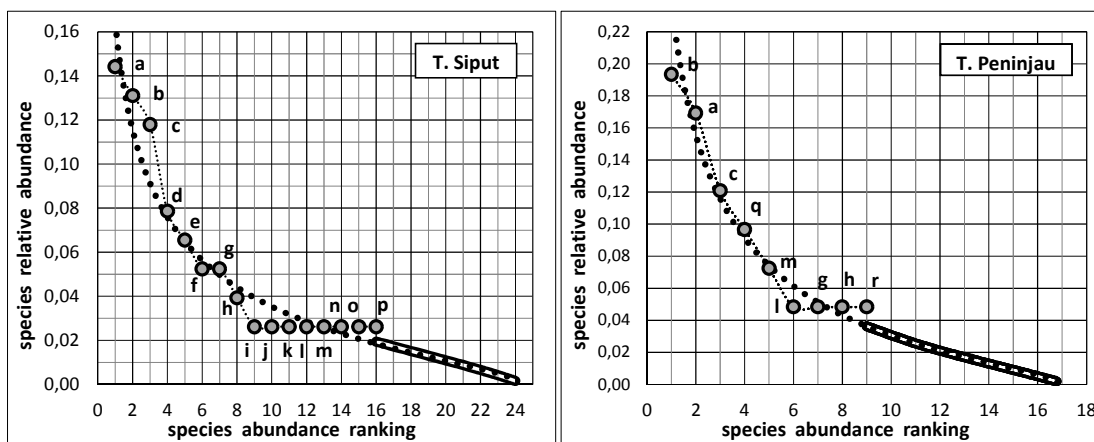


Fig. 11. The two classical models: “log-normal” (coarse dotted line) and “log-series” (fine double line) compared to the Species Abundance Distribution of the community at *Terumbu Peninjau*. Best fit is clearly obtained with the “log-normal” distribution. Note that considering the recorded part of the Species Abundance Distribution *only* would have led to the opposite conclusion: a “J” shaped model such as the “log-series” model would have fit the recorded part best than a “sigmoid” shaped model such as the “log-normal” model



- a: *Linckia laevigata* (Linnaeus 1758) – b: *Linckia multifora* (Lamarck 1816) – c: *Culcita novaeguineae* Müller & Troschel 1842 – d: *Linckia* sp. – e: *Echinaster luzonicus* (Gray 1840) – f: *Fromia monilis* (Perrier 1869) – g: *Linckia guildingi* (Gray 1840) – h: *Leiaster* sp. – i: *Acanthaster planci* (Linnaeus 1758) – j: *Asteropsis carinifera* (Lamarck 1816) – k: *Echinaster callosus* Marrenzeller 1895 – l: *Mithrodia fisheri* Holly 1932 – m: *Fromia indica* (Perrier 1869) – n: *Nordia* sp. – o: *Leiaster speciosus* von Martens 1866 – p: *Ophidiaster hemprichi* Müller & Troschel 1842 – q: *Choriaster granulatus* (Lütken 1869) – r: *Fromia* sp.

Figs. 12 & 13. The same as in Figs. 8 & 9, but, here, the relative abundances of species are left untransformed (instead of being classically log-transformed: [32]) for a better visualization of the good general fit with log-normal model of the recorded part as well. The identities of the eighteen recorded species have been added

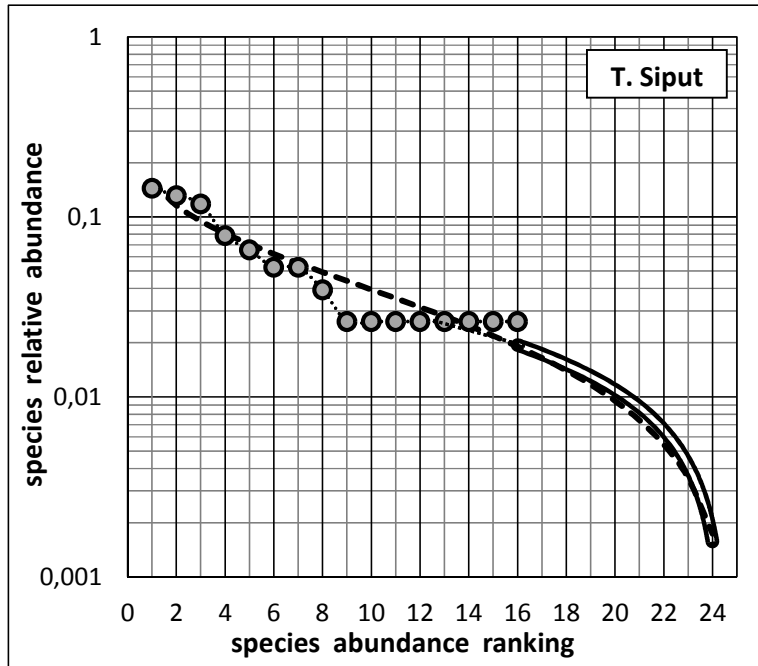


Fig. 14. The species abundance distribution of the community at *Terumbu Siput*, plotted together with the corresponding “broken-stick” distribution (i.e. computed for the same species richness)

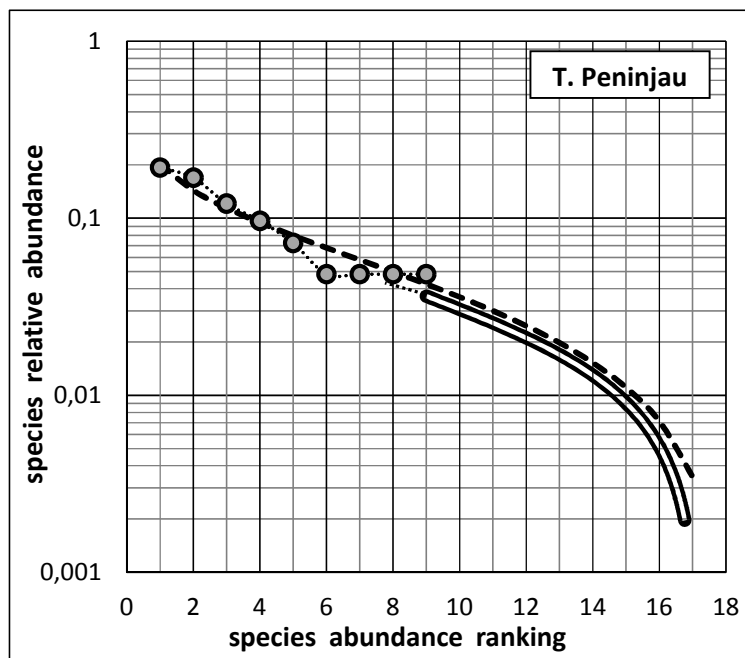


Fig. 15. The species abundance distribution of the community at *Terumbu Peninjau*, plotted together with the corresponding “broken-stick” distribution (i.e. computed for the same species richness)

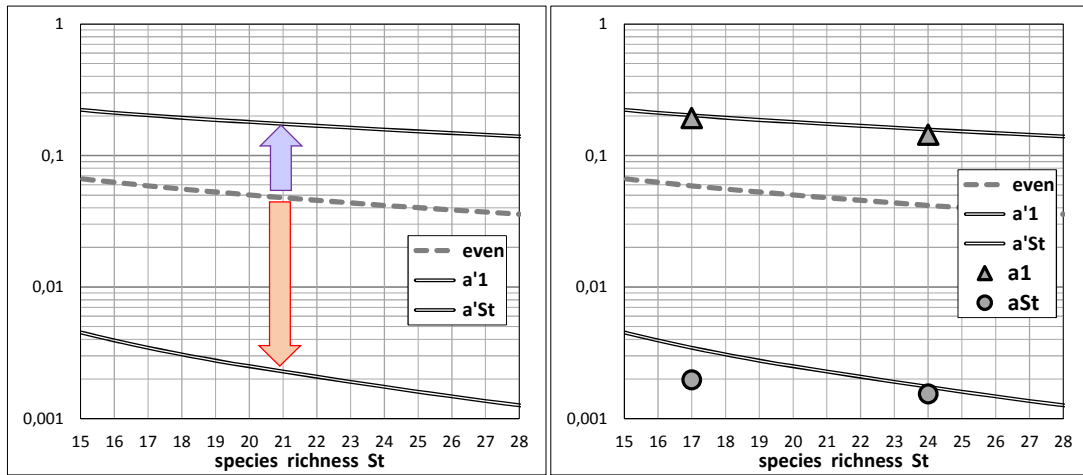
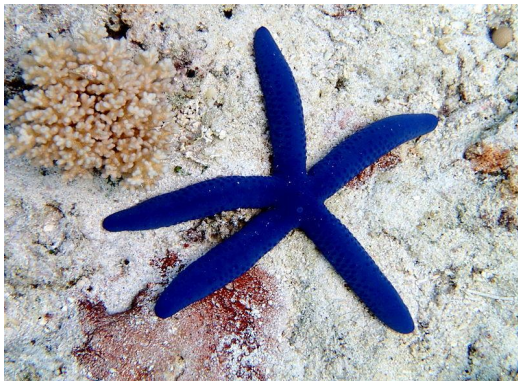
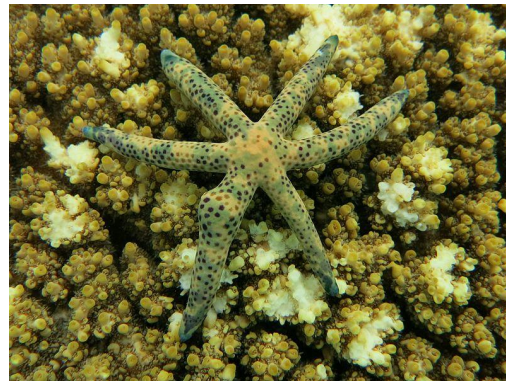


Fig. 16. The variations of the maximum and minimum abundances, a'_1 and a'_{St} of the “broken-stick” model (double lines) and the uniform abundance level ($= 1/S_t$) of the perfectly “even” model (dashed line) as a function of the species richness S_t

Fig. 17. The maximum and minimum abundances, a_1 and a_{St} , for each of the two studied sea-stars communities plotted jointly with the maximum and minimum abundances, a'_1 and a'_{St} of the “broken-stick” model and the uniform abundance level ($= 1/S_t$) of the perfectly “even” model



Linckia laevigata (Linnaeus 1758)
© dr.scott.mills



Linckia multifora (Lamarck 1816)
© Frédéric Ducarme



Culcita novaeguineae Müller & Troschel 1842 © Shizhao



Echinaster luzonicus (Gray 1840)
© Bernard Dupont



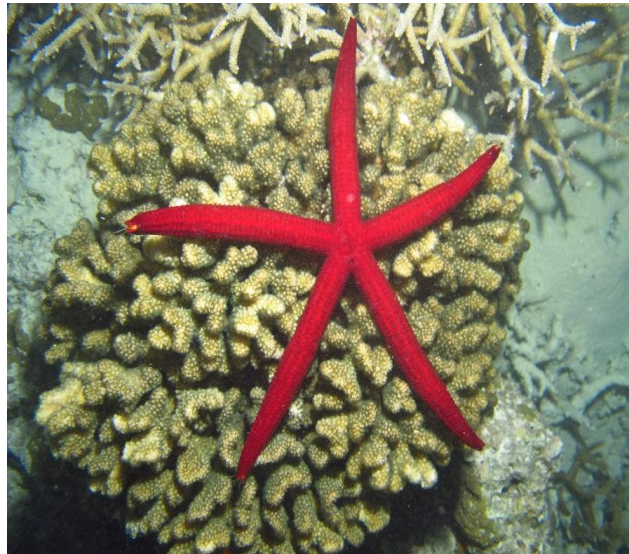
***Fromia monilis* (Perrier 1869)**
© Hectonichus



***Acanthaster planci* (Linnaeus 1758)**
© Michel Dammeron



***Echinaster callosus* Marrenzeller
1895** © Nick Hobgood



***Leiaster speciosus* von Martens 1866**
© MDC Seamark



***Choriaster granulatus* (Lütken, 1869)**
© Ed. Callaghan



***Asteropsis carinifera* (Lamarck 1816)**
© Ph. Bourjon

4. DISCUSSION

Trying to get a comprehensive understanding – species by species – of the internal organization of species-rich communities would normally require long and tedious programs of field investigations, often beyond the usual practical possibilities.

Limiting the scope to the mere evaluation of the total species richness and the overall, synthetic characterization of the hierarchical structuring of abundances will yet still require the *exhaustive* sampling of the studied species assemblage. However, even this less demanding requirement is often difficult to reach in practice, especially when having to deal with species-rich communities including a lot of rare species, as is, for example, often the case with invertebrate faunas.

Fortunately, the implementation of appropriate methods of numerical extrapolation can “force” incomplete samplings and partial inventories to reveal *much more information* than one would have expected *a priori*. Indeed, proper numerical extrapolations of both the Species Accumulation Curve and the Species Abundance Distribution can provide an unexpectedly rich set of additional information relative to those species remaining *undetected* after partial sampling. This, in turn, allows to tackle the main issues relative to the evaluation of true (total) species richness and the hierarchical organization of species abundance, even when communities are only partially sampled – all subjects that otherwise would have required exhaustive inventories.

A thorough analysis of two Sea-Stars communities associated to coral reefs, located in the Central South China Sea, has been conducted accordingly, in compliance with this methodological approach.

4.1 Total Species Richness Estimates and the Forecasted Additional Sampling Efforts Required to Improve Sampling Completeness

At first, the procedure of numerical extrapolation implies selecting the *least-biased* estimator of the number of undetected species. Here, estimators Jackknife-3 and Jackknife-5 are selected for the communities at Trembu Siput and Terembu Peninjau respectively. Accordingly, the total species richness estimated this way reaches 24

species at Trembu Siput and 17 species at Terembu Peninjau (Table 1), which substantially exceeds the recorded numbers (16 and 9 species respectively). This confirms the limited levels of sampling completeness (67% and 53% respectively) and thus justifies, *a posteriori*, the need for implementing numerical extrapolations of these inventories. The importance of selecting for each community the corresponding *least-biased* estimator of the number of undetected species – and the associated least-biased extrapolation of the Species Accumulation Curve – is advocated at Figs. 1 and 3, which both highlight the marked differences that separate the extrapolations of the Species Accumulation Curve associated to different types of estimators. In particular, here, Jackknife-1 and Chao estimators prove being strongly biased negatively (Figs. 1 and 3).

Although the least-biased numerical extrapolation can provide a lot of additional interesting information regarding the set of still unrecorded species, further sampling effort, aiming at increasing the completeness of inventories, might alternatively be considered. In this perspective, reliable forecasts of the additional sampling efforts in order to meet any targeted gain in sampling-completeness would be useful for the optimal planning of the additional efforts to be implemented. The least-biased extrapolation of the Species Accumulation Curve answers appropriately this demand, as shown in Figs. 2 and 4. Clearly, further improvements of sampling completeness would rapidly require very substantial additional efforts. For example, increasing actual completeness at Terembu Siput, from the actual 67% level up to 80%, 90%, 95% completeness levels would require multiplying the actual sample-size ($N_0 = 52$) by a factor 2, a factor 3.5, a factor 7, respectively. Being able to reliably estimate the required additional efforts, as shown above, is of obvious prime interest to rationally decide whether to continue sampling operation any further or to rely only on actual partial inventories, subsequently completed by numerical extrapolation.

4.2 Correction and Extrapolation of the Species Abundance Distribution

As-recorded Species Abundance Distributions need both correction and extrapolation because they are (i) slightly *biased* due to sampling stochasticity and (ii) most importantly, *incomplete*, as no less than eight species had

remained undetected, in each studied community. After correction and extrapolation are applied, the complete development of the Species Abundance Distributions, including the estimated distribution of the abundances of the still undetected species, is made available: Figs. 5, 6, 7. Note that the complementation of Species Abundance Distribution to include the (undetected) less abundant species may be more important than it might seem at first glance. Indeed, less abundant species may possibly have ecological importance no less than more common species, as it has already been repeatedly emphasized by numerous authors [40 – 49]. In addition, considering the *full range* of the Species Abundance Distribution is essential, not only to deliver a full description of the *pattern* of abundances but, also, to question (i) the kind of *process* actually involved in the hierarchical structuring of abundance distribution as well as (ii) the genuine intensity of this structuring process. Indeed, answering these questions imperatively requires comparing the *full range* of Species Abundance Distribution to different theoretical models [5], at the risk, otherwise, to provide severely erroneous inferences.

4.3 Inferring the Type of Process Driving the Hierarchical Structuring of Species Abundances

Considered over their full range, the Species Abundance Distributions of both studied communities clearly fit best the “log-normal” distribution than the “log-series” distribution (Figs. 8 to 11). This suggests that the *process of structuration* of these Sea-Stars communities is likely driven by the combined contributions of many independent factors, rather than by only one (or very few) dominant factor. This, in fact, might well be a rather general trend, as already argued elsewhere [5, 25, 26 – 28]. Moreover, the fairly good fit to the “log-normal” model stands as good within the range of most abundant species (i.e. for lower ranks i). This suggests that, in these Sea-Stars communities, no additional negatively (resp. positively) density-dependent factor actually occurs that, otherwise, would have depressed (resp. increased) the levels of abundance of the more abundant species.

At last, it is also worth noting that relying *only* upon the recorded part of the Species Abundance Distribution would have led to the *opposite conclusion*: indeed, the “J” shape of the “log-series” model fits the *recorded* part best than does the “sigmoid” shape of the “log-normal”

model! This is a new confirmation that relying upon incomplete distributions of abundances *only* (i.e. neglecting the numerical extrapolation when required) may often leads to *erroneous diagnostics*, as already emphasized by several authors [4,6,20,27,50,51].

As regards the recorded species that were, thus, taxonomically identified during limited samplings, the three most common species prove being the same ones in both communities (Figs. 12 and 13). This, indeed, is in expected agreement with the approximate but usually verified correlation between the levels of local abundance and the extent of spatial distribution. By contrast, the identities of those species having lesser abundances largely differ between both communities. Interestingly, however, the abundances distribution of these less abundant species (typically for ranks $i = 4$ to 9) in each of the two communities are, yet, remarkably similar (compare these distributions in Fig. 7). This is a new example of the notion of “functional equivalence despite taxonomical difference” (reviewed in [35]) which is arguably tending to confer some character of resilience among species-rich communities.

4.4 Quantifying the Degree of Hierarchical Structuration of Species Abundances

Here also, considering the full range of the Species Abundance Distribution is necessary, not only to duly include the subset of the still undetected species but, also, to make possible the standardization of the Species Abundance Distribution slope to the corresponding “broken-stick” reference (Figs. 14 & 15). As argued above, this is a key-condition to unveil the *genuine intensity* I_{str} of the process driving the hierarchical structuration of species abundances.

Here (Table 2), the intensity of the structuring process is very close to 1 (i.e. very similar to the intensity in the “broken-stick” model) for the community at Terembu Siput ($I_{str} = 1.003$), while in the community at Terembu Peninjau, the intensity of the structuring process is somewhat (12%) stronger ($I_{str} = 1.128$). In turn, this stronger structuring intensity is only marginally related to the level of dominance of the most abundant species (since the values of $A_1 = (a_1/a'_1)$ for both communities are very similar: Table 2, Fig. 17). Instead, the stronger structuring intensity is mainly due to the lower value of $A_{St} = (a_{St}/a'_{St})$ (0.90 at Terembu Siput and 0.57 only at T. Peninjau: Table 2, Fig. 17). In other words, the

larger intensity of the structuring process highlighted in the community at Terembu Peninjau involves mainly the right-hand part of the abundance distribution that is the subset of less abundant species.

Another interesting question is: how the structuring intensities I_{str} in these two sea-stars communities would compare with the structuring intensities in other types of marine invertebrates communities. The currently available data in this respect remains still limited [21,34,35], but yet suggests that tropical marine gastropod communities tend to be more strongly structured than the two studied Sea-Stars communities.

As regards the *unevenness pattern*, the difference in the degree of unevenness U between the two studied communities proves being still far larger ($> 46\%$, Table 2) than is the difference in the genuine intensity of the structuring process (12%). As already emphasized, such discrepancy between the pattern (U) and the underlying process (I_{str}) is the mere mathematical consequence of the difference in species richness between the two communities.

At last, it should be noted that if the three more abundant species are the same in both communities (labelled a, b, c, in Figs. 12 & 13); the taxonomic composition of the less abundant species is, on the contrary, very different – at least as regards the set of recorded species. Yet, this cannot be considered as firmly conclusive, as it might well be possible that this difference would be less pronounced if the taxonomic identities of the unrecorded species were unveiled. In this respect, numerical extrapolation finds its limit and only further sampling can actually provide a sound answer. An answer, however, at a very substantial extra-cost in term of additional sampling effort. Additional effort that the numerical extrapolation of the Species Accumulation Curves can efficiently help to predict (Figs. 2 & 4).

5. CONCLUSION

When dealing with substantially incomplete species inventories, the *numerical extrapolations* of (i) the Species Accumulation Curve and (ii) the Species Abundance Distribution offer remarkable opportunities to unveil an unexpectedly rich sum of information relative to the set of undetected species. In turn, thanks to the resulting access to the *full range* of the Species Abundance

Distribution, interesting additional information may be derived, regarding the process driving the hierarchical organization of species abundances, in both intensity and mechanism involved. The numerical extrapolations – here applied to the partial inventories of two Sea-Stars communities – demonstrate concretely the wide range of ecological questions that may be addressed and successfully answered, even when no less than one third to almost half of the member-species had remained unrecorded.

In short, this clearly highlights the potential interest of numerical extrapolations applied to partial inventories, in the context of increasingly frequent practice of “quick assessments” of biodiversity, especially when having to deal with highly species-rich assemblages, as is often the case with invertebrate faunas under tropical climates.

ACKNOWLEDGEMENTS

The author appreciates the comments of three anonymous Reviewers and addresses special thanks to S.Y. Kwang, A.T. Shau-Hwai & S. Yasin for the reported field data.

COMPETING INTERESTS

Author has declared that no competing interests exist.

REFERENCES

1. Cam E, Nichols JD, Saue JR, Hines JE. On the estimation of species richness based on the accumulation of previously unrecorded species. *Ecography*. 2002;25: 102-108.
2. Rajakaruna H, Drake DAR, Chan FT, Bailey SA. Optimizing performance of nonparametric species richness estimators under constrained sampling. *Ecology and Evolution*. 2016;6:7311-7322.
3. Chen Y, Shen TJ. Rarefaction and extrapolation of species richness using an area-based Fisher's logseries. *Ecology and Evolution*. 2017;7:10066-10078.
4. Connolly SR, Hughes TP, Bellwood DR. A unified model explains commonness and rarity on coral reefs. *Ecology Letters*. 2017; 20:477-486.
5. May RM. Patterns of species abundance and diversity. In Cody ML, Diamond JM. *Ecology and evolution of communities*. The Belknap Press of Harvard University. 1975; 81-120.

6. McGill BJ, Etienne RS, Gray JS, et al. Species abundance distributions: Moving beyond single prediction theories to integration within an ecological framework. *Ecology Letters*. 2007;10:995-1015.
7. Ulrich W, Ollik M, Uglund KI. A meta-analysis of species-abundance distributions. *Oikos*. 2010;119:1149-1155.
8. Komonen A, Elo M. Ecological response hides behind the species abundance distribution: Community response to low-intensity disturbance in managed grasslands. *Ecology and Evolution*. 2017;7: 8558-8566.
9. Wang X, Ellwood F, Ai D, Zhang R, Wang G. Species abundance distributions as a proxy for the niche-neutrality continuum. *Journal of Plant Ecology*. 2017;rtx 013.
10. Wells JW. Coral reefs: 609-632; In Hedspeth JW editor, *Treatise on marine ecology and paleoecology*. Geological Society of America. Mem. 1957;67.
11. Glynn PW. High complexity food webs in low-diversity eastern Pacific reef-coral communities. *Ecosystems*. 2004;7:358-367.
12. Kwang SY, Shau-Hwai AT, Yasin S. The diversity and abundance of the sea stars (Echinodermata: Asteroidea) from coral reefs of the Central South China Sea. *The Nagisa Westpac Congress*. 2008;25-36.
13. Purwati P, Lane DJW. Asteroidea of the anambas expedition 2002. *The Raffles Bulletin of Zoology*. 2004;(Suppl. no 11): 89-102.
14. Coddington JA, Agnarsson I, Miller JA, Kuntner M, Hormiga G. Undersampling bias: The null hypothesis for singleton species in tropical arthropod surveys. *Journal of Animal Ecology*. 2009;78:573-584.
15. Gotelli NJ, Colwell RK. Estimating species richness. In: *Biological diversity: Frontiers in measurement and assessment*. Magurran AE, McGill BJ (eds.). Oxford University Press, Oxford. 2010;345:39-54.
16. Gotelli NJ, Chao A. Measuring and estimating species richness, species diversity, and biotic similarity from sampling data. In: Levin SA. (ed.) *Encyclopedia of Biodiversity*. Second edition. Waltham, MA: Academic Press. 2013;5:195-211.
17. Béguinot J. Theoretical derivation of a bias-reduced expression for the extrapolation of the species accumulation curve and the associated estimation of total species richness. *Advances in Research*. 2016; 7(3):1-16. DOI:10.9734/AIR/2016/26387
18. Béguinot J. Extrapolation of the species accumulation curve associated to "Chao" estimator of the number of unrecorded species: A mathematically consistent derivation. *Annual Research & Review in Biology*. 2016;11(4):1-19. DOI: 10.9734/ARRB/2016/30522
19. Chao A, Hsieh T, Chazdon RL, Colwell RK, Gotelli NJ. Unveiling the species-rank abundance distribution by generalizing the Good-Turing sample coverage theory. *Ecology*. 2015;96(5):1189-1201.
20. Béguinot J. How to extrapolate species abundance distributions with minimum bias when dealing with incomplete species inventories. *Advances in Research*. 2018; 13(4):1-24. DOI: 10.9734/AIR/2018/39002
21. Béguinot J. Numerical extrapolation of the species abundance distribution unveils the true species richness and the hierarchical structuring of a partially sampled marine gastropod community in the Andaman Islands (India). *Asian Journal of Environment and Ecology*. 2018;6(4):1-23. DOI: 10.9734/AJEE/2018/41293
22. Heip CHR, Herman PMJ, Soetaert K. Indices of diversity and evenness. *Océanis*. 1998;24(4):61-87.
23. Strong WL. Assessing species abundance unevenness within and between plant communities. *Community Ecology*. 2002; 3(2):237-246. DOI: 10.1556/COMEC.3.2002.2.9
24. Grzès IM. Ant species richness and evenness increase along a metal pollution gradient in the Boleslaw zinc smelter area. *Pedobiologia*. 2009;53:65-73.
25. Loreau M. Species abundance patterns and the structure of ground-beetle communities. *Ann. Zool. Fennici*. 1992;28:49-56.
26. Magurran AE, Henderson PA. Explaining the excess of rare species in natural species abundance distributions. *Nature*. 2003;422:714-716.
27. Connolly SR, Hughes TP, Bellwood DR, Karlson RH. Community structure of corals and reef fishes at multiple scales. *Science*. 2005;309:1363-1365.
28. Ulrich W, Soliveres S, Thomas AD, Dougill AJ, Maestre FT. Environmental correlates of species rank-abundance distributions in

- global drylands. *Europe PMC Funders Group*. 2016;20:56-64.
29. Smith B, Wilson JB. A consumer's guide to evenness indices. *Oikos*. 1996;76:70-82.
30. Loiseau N, Gaertner JC. Indices for assessing coral reef fish biodiversity: The need for a change in habits. *Ecology and Evolution*. 2015;5(18):4018-4027.
31. MacDonald ZG, Nielsen SE, Acorn JH. Negative relationships between species richness and evenness render common diversity indices inadequate for assessing long-term trends in butterfly diversity. *Biodiversity Conservation*. 2017;26:617-629.
32. MacArthur RH. On the relative abundance of bird species. *Proceedings of the National Academy of Sciences U.S.A.* 1957;43:293-295.
33. Béguinot J. The hierarchical structuring of species abundances within communities: Disentangling the intensity of the underlying structuring process behind the apparent unevenness pattern. *Advances in Research*. 2018;16(1):1-12.
DOI: 10.9734/AIR/2018/43918
34. Béguinot J. The full hierarchical structuration of species abundances reliably inferred from the numerical extrapolation of still partial samplings: a case study with marine snail communities in Mannar Gulf (India). *Asian Journal of Environment and Ecology*. 2018;7(3):1-27.
DOI: 109734/AJEE/2018/36831
35. Béguinot J. Analyzing the role of environmental stresses on species richness and the process of hierarchical structuring of species abundances in marine Gastropods communities at Suva (Fiji Islands). *International Journal of Environment and Climate Change*. 2018;8(3):200-233.
DOI: 109734/IJECC/2018/44913
36. Pinto-Sanchez NR, Crawford AJ, Wiens JJ. Using historical biogeography to test for community saturation. *Ecology Letters*. 2014;17(9):1077-1085.
37. Cornell HV. Unsaturation and regional influences on species richness in ecological communities: A review of the evidence. *Ecoscience*. 1999;6(3):303-315.
38. Loreau M. Are communities saturated? On the relationship between alpha, beta and gamma diversity. *Ecology Letters*. 2000;3: 73-76.
39. Diamond JM. Assembly of species communities. In Cody ML, Diamond JM. *Ecology and Evolution of Communities*. The Belknap Press of Harvard University. 1975;81-120.
40. Novotny V, Basset Y. Rare species in communities of tropical insect herbivores: Pondering the mystery of singletons. *Oikos*. 2000;89:564-572.
41. Harte J. Tail of death and resurrection. *Nature*. 2003;424:1006-1007.
42. Bracken M, Low N. Realistic losses of rare species disproportionately impact higher trophic levels. *Ecology Letters*. 2012;15: 461-467.
43. Mouillot D, Bellwood DR, Baraloto C, Chave J, Galzin R, Harmelin-Vivien M, Kulbicki M, Lavergne S, Lavorel S, Mouquet N, Paine CET, Renaud J, Thuiller W. Rare species support vulnerable functions in high-diversity ecosystems. *Plos Biol*. 2013;11(5):e1001569.
44. Stuart-Smith RD, Bares AE, Lefcheck JS, Duffy JE, Baker SC, Thomson RJ, Stuart-Smith JF, Hill NA et al. Integrating abundance and functional traits reveals new global hotspots of fish diversity. *Nature*. 2013;501:539-541.
45. Jain M, Flynn DFB, Prager CM, Hart GM, DeVan CM, Ahrestani FS, Palmer MI, Bunker DE, Knops JHM, Jouseau CF, Naeem S. The importance of rare species: A trait-based assessment of rare species contribution to functional diversity and possible ecosystem function in tall-grass prairies. *Ecology and Evolution*. 2014;4(1): 104-112.
46. Henderson PA, Magurran AE. Direct evidence that density-dependent regulation underpins the temporal stability of abundant species in a diverse animal community. *Proceedings of the Royal Society B*. 2014;281.
DOI: 10.1098/rspb.2014.1336
47. Low-Decarie E, Kolber M, Homme P, Lofano A, Dumbrell A, Gonzalez A, Bell G. Community rescue in experimental communities. *Proceedings of the National Academy of Sciences USA*. 2015;112(46): 14307-14312.
48. Leitao RP, Zuanon J, Villéger S, Williams SE, Baraloto C, Fortunel C, Mendonça FP, Mouillot D. Rare species contribute disproportionately to the functional structure of species assemblages. *Proceedings of The Royal Society B*. 2016;283:0084.
DOI: 10.1098/rspb.2016.0084
49. Violle C, Thuillier W, Mouquet N, Munoz F, Kraft NJB, Cadotte MW, Livingstone SW,

- Mouillot D. Functional rarity: The ecology of outliers. *Trends in Ecology*; 2017. DOI: 10.1016/j.tree.2017.02.002
50. Magurran AE. Species abundance distributions: Pattern or process? *Functional Ecology*. 2005;19:177-181.
51. Matthews TJ, Whittaker RJ. On the species abundance distribution in applied ecology and biodiversity management. *Journal of Applied Ecology*. 2015;52:443-454.
52. Béguinot J. An algebraic derivation of Chao's estimator of the number of species in a community highlights the condition allowing Chao to deliver centered estimates. *ISRN Ecology*; 2014. [Article ID 847328]
- DOI: 10.1155/2014/847328
53. Béguinot J. When reasonably stop sampling? How to estimate the gain in newly recorded species according to the degree of supplementary sampling effort. *Annual Research & Review in Biology*. 2015;7(5):300-308. DOI: 10.9734/ARRB/2015/18809
54. O'Hara RB. Species richness estimators: How many species can dance on the head of a pin? *Journal of Animal Ecology*. 2005; 74:375-386.
55. Brose U, Martinez ND, Williams RJ. Estimating species richness: Sensitivity to sample coverage and insensitivity to spatial patterns. *Ecology*. 2003;84(9): 2364-2377.

Appendix 1

Bias-reduced extrapolation of the Species Accumulation Curve and associated bias-reduced estimation of the number of missing species, based on the recorded numbers of species occurring 1 to 5 times

Consider the survey of an assemblage of species of size N_0 (with sampling effort N_0 typically identified either to the number of recorded individuals or to the number of sampled sites, according to the inventory being in terms of either species abundances or species incidences), including $R(N_0)$ species among which f_1, f_2, f_3, f_4, f_5 , of them are recorded 1, 2, 3, 4, 5 times respectively. The following procedure, designed to select the less-biased solution, results from a general mathematical relationship that constrains the theoretical expression of *any* theoretical Species Accumulation Curves $R(N)$: see [17,52,53]:

$$\partial^x R_{(N)}/\partial N^x = (-1)^{(x-1)} f_{x(N)}/C_{N,x} \approx (-1)^{(x-1)} (x!/N^x) f_{x(N)} \quad (\approx \text{as } N \gg x) \quad (\text{A1.1})$$

Compliance with the mathematical constraint (equation (A.1)) warrants *reduced-bias* expression for the extrapolation of the Species Accumulation Curves $R(N)$ (i.e. for $N > N_0$). Below are provided, accordingly, the polynomial solutions $R_x(N)$ that respectively satisfy the mathematical constraint (A1.1), considering increasing orders x of derivation $\partial^x R_{(N)}/\partial N^x$. Each solution $R_x(N)$ is appropriate for a given range of values of f_1 compared to the other numbers f_x , according to Béguinot [17]:

- * for f_1 up to $f_2 \rightarrow R_1(N) = (R(N_0) + f_1) - f_1 \cdot N_0/N$
- * for f_1 up to $2f_2 - f_3 \rightarrow R_2(N) = (R(N_0) + 2f_1 - f_2) - (3f_1 - 2f_2) \cdot N_0/N - (f_2 - f_1) \cdot N_0^2/N^2$
- * for f_1 up to $3f_2 - 3f_3 + f_4 \rightarrow R_3(N) = (R(N_0) + 3f_1 - 3f_2 + f_3) - (6f_1 - 8f_2 + 3f_3) \cdot N_0/N - (-4f_1 + 7f_2 - 3f_3) \cdot N_0^2/N^2 - (f_1 - 2f_2 + f_3) \cdot N_0^3/N^3$
- * for f_1 up to $4f_2 - 6f_3 + 4f_4 - f_5 \rightarrow R_4(N) = (R(N_0) + 4f_1 - 6f_2 + 4f_3 - f_4) - (10f_1 - 20f_2 + 15f_3 - 4f_4) \cdot N_0/N - (-10f_1 + 25f_2 - 21f_3 + 6f_4) \cdot N_0^2/N^2 - (5f_1 - 14f_2 + 13f_3 - 4f_4) \cdot N_0^3/N^3 - (-f_1 + 3f_2 - 3f_3 + f_4) \cdot N_0^4/N^4$
- * for f_1 larger than $4f_2 - 6f_3 + 4f_4 - f_5 \rightarrow R_5(N) = (R(N_0) + 5f_1 - 10f_2 + 10f_3 - 5f_4 + f_5) - (15f_1 - 40f_2 + 45f_3 - 24f_4 + 5f_5) \cdot N_0/N - (-20f_1 + 65f_2 - 81f_3 + 46f_4 - 10f_5) \cdot N_0^2/N^2 - (15f_1 - 54f_2 + 73f_3 - 44f_4 + 10f_5) \cdot N_0^3/N^3 - (-6f_1 + 23f_2 - 33f_3 + 21f_4 - 5f_5) \cdot N_0^4/N^4 - (f_1 - 4f_2 + 6f_3 - 4f_4 + f_5) \cdot N_0^5/N^5$

The associated non-parametric estimators of the number Δ_J of missing species in the sample [with $\Delta_J = R(N = \infty) - R(N_0)$] are derived immediately:

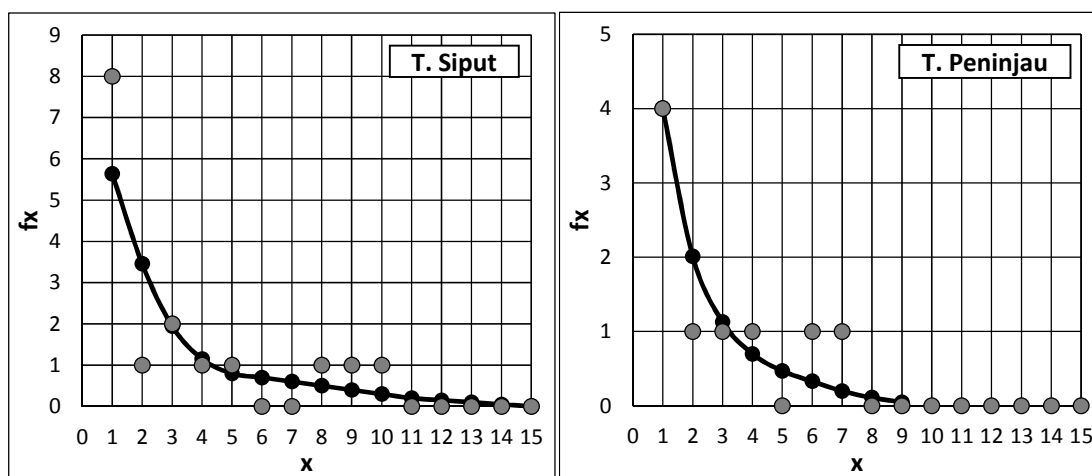
- * $0.6 f_2 < f_1 \leq f_2 \rightarrow \Delta_{J1} = f_1 ; R_1(N)$
- * $f_2 < f_1 \leq 2f_2 - f_3 \rightarrow \Delta_{J2} = 2f_1 - f_2 ; R_2(N)$
- * $2f_2 - f_3 < f_1 \leq 3f_2 - 3f_3 + f_4 \rightarrow \Delta_{J3} = 3f_1 - 3f_2 + f_3 ; R_3(N)$
- * $3f_2 - 3f_3 + f_4 < f_1 \leq 4f_2 - 6f_3 + 4f_4 - f_5 \rightarrow \Delta_{J4} = 4f_1 - 6f_2 + 4f_3 - f_4 ; R_4(N)$
- * $f_1 > 4f_2 - 6f_3 + 4f_4 - f_5 \rightarrow \Delta_{J5} = 5f_1 - 10f_2 + 10f_3 - 5f_4 + f_5 ; R_5(N)$

N.B. 1: As indicated above (and demonstrated in details in Béguinot [17], this series of inequalities define the ranges that are best appropriate, respectively, to the use of each of the five estimators, JK-1 to JK-5. That is the respective ranges within which each estimator will benefit of minimal bias for the predicted number of missing species.

Besides, it is easy to verify that another consequence of these preferred ranges is that the selected estimator will *always* provide the highest estimate, as compared to the other estimators. Interestingly, this mathematical consequence, of general relevance, is in line with the already admitted opinion that all non-parametric estimators provide *under*-estimates of the true number of missing species [2,3,15, 16,54]. Also, this shows that the approach initially proposed by BROSE et al. [55] – which has regrettably suffered from its somewhat difficult implementation in practice – might be advantageously reconsidered, now, in light of the very simple selection key above, of *far much easier practical use*.

N.B. 2: In order to reduce the influence of drawing stochasticity on the values of the f_x , the as-recorded distribution of the f_x should preferably be smoothened: this may be obtained either by rarefaction processing or by regression of the as-recorded distribution of the f_x versus x .

N.B. 3: For f_1 falling beneath $0.6 \times f_2$ (that is when sampling completeness closely approaches exhaustivity), then Chao estimator may be selected: see reference [18].



Figs. A1 & A2. The recorded values of the numbers f_x of species recorded x -times (grey discs) and the regressed values of f_x (black discs) derived to reduce the consequence of stochastic dispersion during sampling

Appendix 2

Correction and extrapolation of the as-recorded Species Abundance Distribution (S.A.D.)

N.B.: details regarding the derivation of the following expressions are provided in [20].

1) Correction for bias of the recorded part of the S.A.D.

The bias-corrected expression of the true abundance, \tilde{a}_i , of species of rank 'i' in the S.A.D. is given by:

$$\tilde{a}_i = p_i \cdot (1 + 1/n_i) \cdot (1 - f_1/N_0) / (1 + R_0/N_0) \quad (\text{A2.1})$$

where N_0 is the actually achieved sample size, $R_0 (=R(N_0))$ the number of recorded species, among which a number f_1 are singletons (species recorded only once), n_i is the number of recorded individuals of species 'i', so that $p_i = n_i/N_0$ is the recorded frequency of occurrence of species 'i', in the sample. The crude recorded part of the "S.A.D." – expressed in terms of the series of as-recorded frequencies $p_i = n_i/N_0$ – should then be replaced by the corresponding series of expected true abundances, \tilde{a}_i , according to equation (A2.1).

2) Extrapolation of the recorded part of the S.A.D. accounting for the complementary abundance distribution of the set of unrecorded species

The following expression stands for the estimated abundance, a_i , of the unrecorded species of rank i (thus for $i > R_0$):

$$a_i = (2/N_i) \cdot (1 - [\partial R(N)/\partial N]_{N_i}) / (1 + R(N_i)/N_i) \quad (\text{A2.2})$$

which, in practice, comes down to:

$$a_i \approx (2/N_i) / (1 + R(N_i)/N_i) \quad (\text{A2.3})$$

as $f_1(N)$ already becomes quite negligible as compared to N for the extrapolated part.

This equation provides the extrapolated distribution of the species abundances a_i (for $i > R(N_0)$) as a function of the least-biased expression for the extrapolation of the species accumulation curve $R(N)$ (for $N > N_0$), 'i' being equal to $R(N_i)$. The key to select the least-biased expression of $R(N)$ is provided at Appendix 1.

Appendix 3

The trivial contribution of the level of species richness to the degree of structuring of species abundances

All things equal otherwise, the larger the species richness, the weaker is the slope of the Species Abundance Distribution. This can be easily exemplified and quantified, on a theoretical basis, by considering a theoretically constant structuring process - such as the random distribution of the relative abundances that characterises the “broken-stick” distribution model. By applying this model successively to a series of communities with increasing species richness, a steadily decrease of the slope of abundance distributions is highlighted: Fig. A3

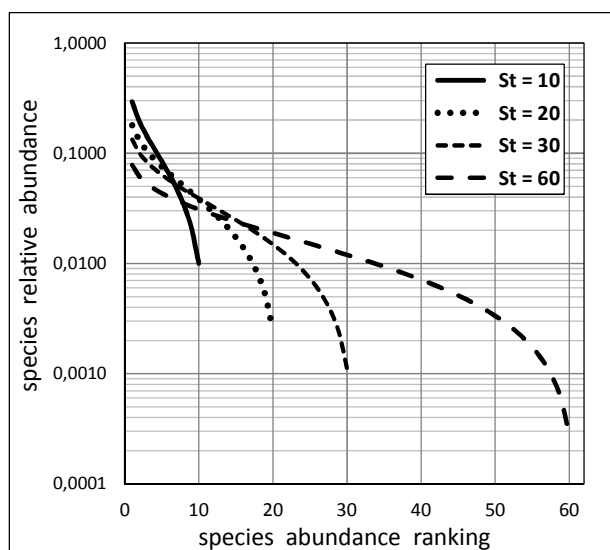


Fig. A3. The “broken-stick” distribution model applied to species communities with increasing species richness $S_t = 10, 20, 30, 60$. Although the theoretical structuring process involved in the “broken-stick” model remains unchanged (random apportionment of relative abundances among member species), the slope of the species abundance distribution strongly depends upon (and monotonously decreases with) the level of species richness S_t

© 2018 Béguinot; This is an Open Access article distributed under the terms of the Creative Commons Attribution License (<http://creativecommons.org/licenses/by/4.0>), which permits unrestricted use, distribution, and reproduction in any medium, provided the original work is properly cited.

Peer-review history:

The peer review history for this paper can be accessed here:
<http://www.sdiarticle3.com/review-history/46272>